

Affinity of extracellular phosphatases for ELF97 phosphate in aquatic environments

Jiří Nedoma^{A,D}, France Van Wambeke^B, Alena Štrojsová^{A,C},
Martina Štrojsová^{A,C} and Solange Duhamel^B

^ABiological Centre of the Academy of Sciences of the Czech Republic, v.v.i., Hydrobiological Institute, Na sádkách 7, 37005 České Budějovice, Czech Republic.

^BLaboratoire de Microbiologie Géochimie et Ecologie Marines, CNRS-UMR 6117, Campus de Luminy, case 901, 13 288 Marseille cedex 9, France.

^CFaculty of Biological Sciences, University of South Bohemia, Branišovská 31, 370 05 České Budějovice, Czech Republic.

^DCorresponding author. Email: nedoma@hbu.cas.cz

Abstract. Recently, the phosphatase substrate ELF97 phosphate (ELFP) has been employed to study the presence of extracellular phosphatases in different plankton populations in natural aquatic environments. Kinetic properties of ELFP hydrolysis by natural extracellular phosphatases are, however, mostly unknown. We indirectly studied the affinity of extracellular phosphatases for ELFP in different aquatic environments through its ability to inhibit the hydrolysis of 4-methylumbelliferyl phosphate (4MUP). Values of inhibition constants, K_i , which correspond to the concentrations necessary for half saturation of phosphatases by ELFP, were lowest ($0.18\text{--}4.5\ \mu\text{mol L}^{-1}$) in the oligotrophic Mediterranean Sea. We found higher values (i.e. lower affinity) in oligo- to mesotrophic acidified lakes ($5.2\text{--}14\ \mu\text{mol L}^{-1}$), in a eutrophic reservoir ($13\text{--}35\ \mu\text{mol L}^{-1}$) and in a pure culture of the marine bacterium *Alteromonas infernus* ($29\ \mu\text{mol L}^{-1}$). ELFP had a pronounced effect on the parameter K_M (Michaelis constant) of 4MUP saturation kinetics, while its effect on the parameter V_{\max} was low. This behaviour is compatible with the assumption of competitive interaction between 4MUP and ELFP. Our experiments indicated that the assay ELFP concentration in the detection kit used was $250\text{--}500\ \mu\text{mol L}^{-1}$ (after the recommended dilution to a ratio of 1:20), which would ensure >99% saturation of extracellular phosphatases in marine environments and >90% saturation in the studied fresh waters.

Additional keywords: freshwater, marine.

Introduction

Phosphorus availability is considered to be a factor that limits productivity in many freshwater and certain marine systems (e.g. Schindler 1977; Elser *et al.* 1995; Van Wambeke *et al.* 2002). During periods of phosphorus limitation, concentrations of dissolved inorganic phosphorus, the only directly bioavailable phosphorus form, drop regularly to negligible values (Taylor and Lean 1991; Moutin *et al.* 2002). Synthesis of inducible extracellular phosphatases is a common adaptive response of microbial cells to phosphorus shortage (Gage and Gorham 1985; Jansson *et al.* 1988; Hoppe 2003); however, phosphorus stress might not be the only motive for their synthesis (Hoppe 2003) and non-microbial origin of phosphatases may be important in particular cases (Jean *et al.* 2003). Phosphatases cleave molecules of organic esters of phosphoric acid into orthophosphate and organic parts, thus converting the mostly non-bioavailable dissolved organic phosphorus to inorganic phosphorus. Recently, the phosphatase substrate ELF97 phosphate (ELFP) was employed to study the presence of extracellular phosphatases in different plankton populations in natural aquatic environments (González-Gil *et al.* 1998; Rengefors *et al.* 2001; Štrojsová *et al.* 2003, 2005).

The product of ELFP hydrolysis, a brightly fluorescent ELF97 alcohol (ELFA), is water-insoluble and precipitates at the site of its origin, thus labelling cells with phosphatase activity.

Very little is known about the kinetics of ELFP hydrolysis by extracellular phosphatases in natural aquatic systems. The ELFP hydrolysis is only an initial step of a complicated process leading finally (and not necessarily) to the formation of observable ELFA precipitates, and these two processes may be decoupled. Moreover, the quantification of the fluorescence of ELFA precipitates is hindered by the low ELFA fluorescence yield and by the complicated methodology. The affinity of extracellular phosphatases for ELFP is therefore not known and, as a consequence, neither is the degree of phosphatase saturation by ELFP in most experiments. Experiments with conventional substrates like 4MUP suggest a great variability in substrate affinity in different aquatic systems (e.g. differences in K_M of three orders of magnitude for 4MUP among lakes, Nedoma *et al.* 2003a).

Most frequently, ELFP is applied as the part of a special 'endogenous phosphatase detection kit', developed originally for use in histochemistry by Molecular Probes (Cox and Singer 1999). However, the concentration of ELFP used in the kit is

Table 1. Basic parameters of the samples used in this study, and kinetic parameters (\pm s.d.) K_M and V_{max} for 4-methylubelliferyl phosphate (4MUP) hydrolysis and inhibition constants K_i for inhibition of 4MUP hydrolysis by ELF97 phosphate (ELFP)

4MUP range = range of 4MUP concentrations used in saturation experiments; K_i = inhibition constant determined in displacement experiments using pure ELFP; NWM = north-west Mediterranean; b.d.l. = below the detection limit of 30 nmol L⁻¹; n.a. = not available

Date	Locality (source)	Temp °C	pH	SRP nmol L ⁻¹	V_{max} nmol L ⁻¹ h ⁻¹	4MUP K_M μ mol L ⁻¹	4MUP range μ mol L ⁻¹	ELFP K_i μ mol L ⁻¹
8 Oct 04	DYFAMED station (NWM)	21	8.3	35	2.6 \pm 0.1	0.076 \pm 0.011	0.01–2	0.44 \pm 0.12
10 Oct 04	DYFAMED station	20	8.3	35	2.7 \pm 0.1	0.069 \pm 0.009	0.01–2	0.55 \pm 0.22
15 Oct 04	DYFAMED station	19	8.3	35	2.7 \pm 0.1	0.082 \pm 0.008	0.01–2	0.18 \pm 0.04
26 May 05	SOFCOM station (Marseille Bay)	17	8.3	30	31 \pm 2	0.41 \pm 0.09	0.2–40	4.5 \pm 1.7
1 Feb 05	Plešné Lake (Czech Republic)	0	4.9	b.d.l.	1600 \pm 100	1.3 \pm 0.4	0.2–20	5.2 \pm 0.4
22 Feb 05	Čertovo Lake (Czech Republic)	0	4.4	35	700 \pm 50	2.9 \pm 0.7	0.2–20	14 \pm 2
14 Mar 05	Čertovo Lake	0	4.5	30	260 \pm 20	1.4 \pm 0.6	0.2–40	12 \pm 5
18 May 05	Čertovo Lake	9.5	4.6	b.d.l.	270 \pm 20	1.6 \pm 0.5	0.2–40	13 \pm 2
13 Sep 05	Římov Reservoir (Czech Republic)	19	8.0	115	53 \pm 4	1.2 \pm 0.3	0.025–20	16 \pm 5
22 Sep 05	Římov Reservoir	14.5	6.9	95	28 \pm 2	1.2 \pm 0.3	0.025–20	23 \pm 10
4 Oct 05	Římov Reservoir	14	7.1	220	31 \pm 11	5.7 \pm 5.1	0.025–20	35 \pm 4
26 May 05	Bacterial culture <i>Alteromonas infernus</i>	20	7.5	n.a.	417 \pm 5	0.23 \pm 0.02	2–100	29 \pm 5

not known, which limits the interpretation of results. Pure ELFP of a known concentration was successfully employed in some studies as an alternative to the kit ELFP (Nedoma *et al.* 2003b; Štrojsová *et al.* 2003, 2005).

One possible way to obtain information about the kinetics of ELFP hydrolysis by naturally occurring extracellular phosphatases is to study the kinetics indirectly, studying the interaction of ELFP with the hydrolysis of another substrate that is readily measurable, e.g. 4MUP, and by employing the theoretical background developed for studying enzyme inhibition (Dixon and Webb 1964). Within the framework of this approach, 4MUP acts as a substrate and ELFP as an inhibitor. Analysis of the inhibition kinetics enables the estimation of the inhibition constant, K_i , which is a measure of enzyme affinity for the inhibitor. Thus, we may indirectly estimate phosphatase affinity for ELFP, as well as the concentration of ELFP in the endogenous phosphatase detection kit, by comparing its efficiency in the inhibition of 4MUP hydrolysis with the efficiency of pure ELFP of known concentration used as a standard. Another intriguing question, not yet answered (or raised), is whether the enzymes that hydrolyse ELFP and 4MUP (as well as other conventional substrates) in natural environments are identical. Implicitly, a positive answer is assumed but there is little experimental evidence on the subject (Štrojsová *et al.* 2003).

The specific aims of the experiments described in the present study were to indirectly determine through inhibition of 4MUP hydrolysis (1) the affinity of extracellular phosphatases in different aquatic environments towards ELFP, and (2) whether the same enzymes hydrolyse ELFP and 4MUP.

Materials and methods

Samples

We used water samples from different marine and freshwater environments taken from 5-m (marine) and 0.5-m depth (freshwater). Additionally, a pure bacterial culture was also examined. Some relevant characteristics of samples are given

in Table 1. The Dynamique des Flux de Matière en Méditerranée (DYFAMED) station is situated in the Western Mediterranean Sea, ~52 kilometres south of Nice (43°25'N, 07°52'E; depth 2350 m). A nitrate-to-phosphate ratio of ~20 in deep waters and thermal stratification in summer and autumn creates a situation of strong phosphorus limitation (Marty *et al.* 2002). The Service d'Observation du Centre d'Océanologie (SOFCOM) station is located in the southern part of Bay of Marseille, France (43°18'N, 5°24'E; depth 30 m) and it is under greater anthropogenic influence.

The Lakes Čertovo and Plešné are small, acidified mountain lakes (maximum depths 34 and 17 m respectively) located in the Bohemian Forest, Czech Republic (30°12'E, 49°10'N) at altitudes of 1040–1080 m. Lakes are permanently phosphorus-depleted owing to aluminium-phosphorus co-precipitation, which removes it from the water column (Kopáček *et al.* 2000). The eutrophic Římov reservoir is located on the river Malše in South Bohemia (49°10'N, 15°46'E; maximum depth 43 m, altitude 470 m), Czech Republic. Phosphorus limitation develops regularly in the epilimnion in spring and summer (Vrba *et al.* 1993, 1995; Štrojsová *et al.* 2003).

The marine bacterial strain *Alteromonas infernus* (γ -proteobacteria), possessing inducible extracellular phosphatases, was cultivated in batch conditions in an artificial seawater medium that contained pyruvate (0.33 mmol L⁻¹ C) and orthophosphate (0.825 μ mol L⁻¹ P) as single carbon and phosphorus sources respectively. We conducted experiments during the stationary phase of culture growth after strong phosphorus limitation had developed.

Chemicals

We purchased ELF97 phosphate (ELFP) from the producer, Molecular Probes (Eugene, OR, USA), either as a pure substance (5 mmol L⁻¹ water solution, 0.2- μ m filtered, Catalogue No. E-6588), abbreviated as ELFP^{PURE} hereafter, or as a part of the endogenous phosphatase detection kit (Catalogue No. E-6601, Component A, Phosphatase substrate, 20 \times concentrate,

concentration not specified, in 2 mmol L^{-1} sodium azide), abbreviated hereafter as ELFP^{KIT}. We stored the original (supplied) solutions of ELFP^{PURE} and ELFP^{KIT} at -20°C and $+4^{\circ}\text{C}$, respectively, and filtered them through a spin filter of $0.2\text{-}\mu\text{m}$ porosity just before use. The 4-methylumbelliferyl phosphate disodium salt (4MUP) and 4-methylumbelliferone (4MU) were purchased from Sigma (www.sigma.com).

Extracellular phosphatase activity

We fluorimetrically determined extracellular phosphatase activity using 4MUP as a substrate (Hoppe 1983) and 4MU as a standard. We performed incubations in duplicates, in the dark and at *in situ* temperature, adding $25\text{ }\mu\text{L}$ of the substrate (4MUP) and 25 or $75\text{ }\mu\text{L}$ of the inhibitor (ELFP diluted in H_2O , replaced by H_2O in controls) to 2.5 mL of water sample. In parallel, we conducted calibrations relating fluorescence to 4MU concentration in each sample. Owing to specific requirements imposed by the character of samples (pH, enzyme activity) and the laboratory equipment available, the details of the method slightly differed for marine (Van Wambeke *et al.* 2002) and freshwater (Vrba *et al.* 1995) samples and for the bacterial culture. For marine samples, we measured fluorescence at *in situ* pH (8.3) for up to 6 h. For freshwater samples, incubations lasted from 15 to 240 min at *in situ* pH, and were completed by the addition of $100\text{ }\mu\text{L}$ of 0.2 mol L^{-1} HgCl (present from the beginning in blanks). We alkalisied samples by the addition of $100\text{ }\mu\text{L}$ of the mixture of 2 mol L^{-1} NaOH + 0.2 mol L^{-1} EDTA (final pH > 11) before fluorescence measurement. In bacterial cultures (buffered to pH 7.5 with 10 mmol L^{-1} TRIS/HCl), we read fluorescence throughout 2-h incubations.

Inhibition experiments

We performed two types of inhibition experiments: 'displacement experiments' and 'saturation experiments'. We use these terms throughout the present study to distinguish two different experimental designs: (1) displacement experiments, in which we used different ELFP concentrations and a single 4MUP concentration (Fig. 1a), and (2) saturation experiments, in which we used a single ELFP concentration and different 4MUP concentrations (Fig. 1b). In each inhibition experiment, both ELFP^{PURE} and ELFP^{KIT} were applied in parallel at the same dilution (sets of dilutions) of the original solutions delivered by the manufacturer. The ELFP concentration referred to e.g. as 10^{-3} therefore means a final assay ELFP concentration that is 1000-times lower than that of the original solution delivered. In the case of ELFP^{PURE}, supplied as 5 mmol L^{-1} solution, this corresponds to $5\text{ }\mu\text{mol L}^{-1}$. Consequently, the constants K_i^{PURE} and K_i^{KIT} , characterising affinity of ELFP^{PURE} and ELFP^{KIT} to phosphatases, are given in units of ELFP dilution as well. In the case of K_i^{PURE} , the values can be recalculated to molar units as above.

Displacement experiments

In inhibition experiments of the displacement type, we examined the effects of ELFP, applied in a series of concentrations, on the hydrolysis velocity of a substrate (i.e. 4MUP) applied at a single low concentration (Fig. 1a). We kept 4MUP concentration, S , as low as possible, preferably $S < K_M$ (K_M is the Michaelis constant for the 4MUP hydrolysis by extracellular phosphatases present

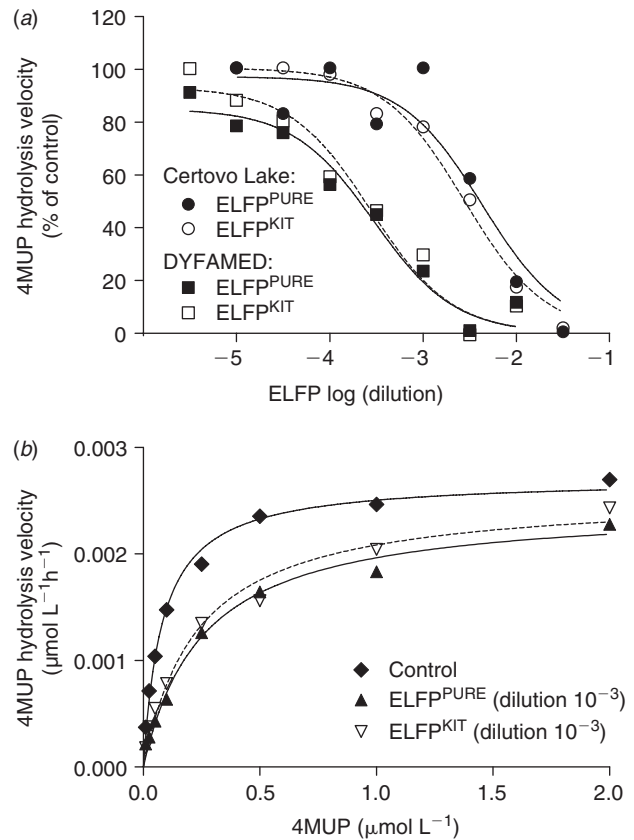


Fig. 1. (a) Examples of displacement curves: inhibition of hydrolysis of 4-methylumbelliferyl phosphate (4MUP) by ELF97 phosphate, applied at different concentrations, in samples taken at the DYFAMED station (north-west Mediterranean) on 10 October 2004, and in Lake Čertovo on 14 March 2005. Final concentration of 4MUP was $0.1\text{ }\mu\text{mol L}^{-1}$ in a marine sample and $1\text{ }\mu\text{mol L}^{-1}$ in a lake sample. ELFP concentrations are given as logarithm of dilution of the original solutions purchased from the manufacturer (Molecular Probes), either as a part of the endogenous phosphatase detection kit (ELFP^{KIT}; open symbols) or as a pure substance (ELFP^{PURE}; closed symbols). (b) An example of a saturation experiment: a comparison of effects of ELFP^{KIT} and ELFP^{PURE} (applied at the same dilution of 10^{-3}) on saturation of 4MUP hydrolysis in the sample taken from the DYFAMED station on 4 October 2004.

in the sample examined, determined in a preceding experiment; Table 1). Final ELFP dilutions ranged from 3×10^{-6} to 3×10^{-2} (Fig. 1a). To determine IC_{50} , which corresponds to the ELFP concentration (dilution) that causes 50% inhibition of 4MUP hydrolysis, we expressed the 4MUP hydrolysis velocities occurring in the presence of different ELFP concentrations (dilutions) as percentages of the control values occurring in the absence of ELFP, and plotted them in a semi-logarithmic way as a function of ELFP dilution (Fig. 1a). We calculated IC_{50} using non-linear regression by fitting data with a decelerating three-parameter sigmoid function (displacement curve), assuming competitive inhibition

$$v = \text{Min} + (\text{Max} - \text{Min})(1 + I/IC_{50})^{-1} \quad (1)$$

where v is the substrate (4MUP) hydrolysis velocity (% of control), Max and Min are the upper and lower plateaus of

the displacement curve (% of control), respectively, and I is inhibitor concentration (i.e. final ELFP dilution, dimensionless). The lower plateau (Min) of the displacement curve was set to zero, because we usually had not enough data in the range of high ELFP concentrations with which to estimate it reliably (cf. Fig. 1a). Because IC_{50} values do not directly correspond to K_i unless the substrate concentration used is substantially lower than K_M for this substrate (i.e. $S < 0.1K_M$: a condition not fulfilled in our experiments), we corrected the IC_{50} values according to Chang and Prusoff (1973) to obtain inhibition constants K_i

$$K_i = IC_{50}(1 + S/K_M)^{-1} \quad (2)$$

where S is the substrate (4MUP) concentration used in the displacement experiment, and K_M is the Michaelis constant for 4MUP hydrolysis determined in parallel saturation experiments.

Saturation experiments

In inhibition experiments of the saturation type, we studied the effect of ELFP on the kinetic parameters K_M and V_{max} of 4MUP hydrolysis. We measured 4MUP hydrolysis using a series of 4MUP concentrations in the absence (control) and in the presence of a single concentration of ELFP. In each saturation experiment, we tried to choose an inhibitor concentration, I , slightly above K_i of ELFP inhibition of 4MUP hydrolysis (determined in parallel inhibition experiments), preferably $K_i < I < 2K_i$. We determined the parameters V_{max} (maximum hydrolysis velocity) and K_M (Michaelis constant, which reflects enzyme affinity for 4MUP) using non-linear regression by fitting the data with the Michaelis–Menten equation

$$v = V_{max} \times S/(S + K_M)^{-1} \quad (3)$$

where v is 4MUP hydrolysis velocity ($\mu\text{mol L}^{-1} \text{h}^{-1}$) and S is 4MUP concentration. To obtain K_i , the inhibition constant of 4MUP hydrolysis inhibition by ELFP, we used the equation assuming competitive inhibition

$$K_i = I(1 + K_M^{\text{CONTROL}}/K_M^{\text{ELFP}})^{-1} \quad (4)$$

where K_M^{CONTROL} is the Michaelis constant determined in the absence of the inhibitor, and K_M^{ELFP} is that determined in the presence of ELFP. The inhibition constants K_i in Eqns 2 and 4 are identical. Consequently, their estimates from displacement and saturation experiments in Eqns 2 and 4, respectively, should theoretically give the same value. The constant K_i numerically corresponds to the concentration of inhibitor necessary for 50% saturation of enzymes in the absence of substrate, and can be thus interpreted as a measure of enzyme affinity towards the inhibitor, similarly to K_M in the case of a substrate.

Statistical treatment

The Prism 4 (GraphPad Software, San Diego, USA) statistical package was used to perform non-linear regressions and to test differences between parameter estimates obtained in experiments with inhibitors or control (using the F -test).

Results

The basic hydrological variables and kinetic properties of extracellular phosphatases differed considerably among the samples examined in the present study (Table 1). The soluble reactive phosphorus concentrations varied from the values below detection limit to 220 nmol L^{-1} , pH ranged from 4.5 to 8.0, and temperature ranged from 0 to 21°C . The maximum velocity of 4MUP hydrolysis, V_{max} , varied within nearly three orders of magnitude ($2.6\text{--}1600 \text{ nmol L}^{-1} \text{h}^{-1}$), and the Michaelis constant, K_M , varied by two orders of magnitude ($0.076\text{--}6.3 \mu\text{mol L}^{-1}$). The lowest values of both parameters were found in marine samples, whereas the highest values occurred in a eutrophic reservoir (Table 1).

In displacement experiments, measured data on inhibition of 4MUP hydrolysis (Fig. 1a) fitted well the competitive inhibition model (Eqn 1). However, in some experiments, the lower plateaus of displacement curves were not sufficiently completed with measured data because of low phosphatase affinity for ELFP (i.e. high IC_{50} , the ELFP concentration causing 50% inhibition). Theoretically, the inhibitor concentration has to surpass IC_{50} by at least two orders of magnitude in order to lead to complete inhibition ($>99\%$; cf. Eqn 1). In contrast, the highest ELFP concentration applied in the present study was substantially (8- to 124-times) higher than IC_{50} in only 7 of our 12 experiments. In these experiments, the maximum inhibition ranged from 80 to 100% (average 92%), which was not statistically different from the expected 88 to 99% inhibition (average 93%) calculated from Eqn 1 (paired t -test, $P = 0.73$). In the 5 remaining displacement experiments, we observed inhibition only up to 54–81% owing to the small difference (1.1- to 5.5-fold) between the highest ELFP concentration applied and IC_{50} .

In saturation experiments, the presence of ELFP significantly decreased the apparent K_M of 4MUP hydrolysis compared with the control without ELFP in all cases (F -test; $P < 0.05$), while the effect on V_{max} was statistically significant in 2 of 6 cases only (increase to 120 and 340% of control value) (F -test; $P < 0.05$). The competitive inhibition model expects no effect on V_{max} and an inhibitor-concentration-dependent (cf. Eqn 4) decrease in K_M .

We found the highest affinity of extracellular phosphatases for ELFP (i.e. the lowest K_i^{PURE} ; Table 1) in marine samples ($0.18\text{--}4.5 \mu\text{mol L}^{-1}$), followed by acidified lakes ($5.2\text{--}11 \mu\text{mol L}^{-1}$) and the freshwater reservoir ($16\text{--}34 \mu\text{mol L}^{-1}$). The K_M values of 4MUP hydrolysis in the aquatic systems studied followed the same order as the corresponding K_i values, the latter being on average 8 times higher (excluding the bacterial culture with a difference of two orders of magnitude; Table 1).

Both $\text{ELFP}^{\text{PURE}}$ and ELFP^{KIT} inhibited 4MUP hydrolysis, typically with similar efficiency (see Fig. 1 for examples of raw data). In order to compare inhibition efficiencies of $\text{ELFP}^{\text{PURE}}$ and ELFP^{KIT} quantitatively, we determined their respective inhibition constants, K_i^{PURE} and K_i^{KIT} (using Eqn 2 or Eqn 4), and calculated ratios of these constants ($K_i\text{-ratio} = K_i^{\text{KIT}}/K_i^{\text{PURE}}$), for each experiment. Fig. 2 shows comparisons of K_i^{PURE} and K_i^{KIT} values estimated in displacement (Fig. 2a) and saturation (Fig. 2b) experiments. Differences between K_i^{PURE} and K_i^{KIT} were statistically significant ($P < 0.05$, F -test) in 3 of the total of 12 displacement experiments and in 1 of 7 saturation experiments performed: $\text{ELFP}^{\text{PURE}}$ was more effective in 1 case

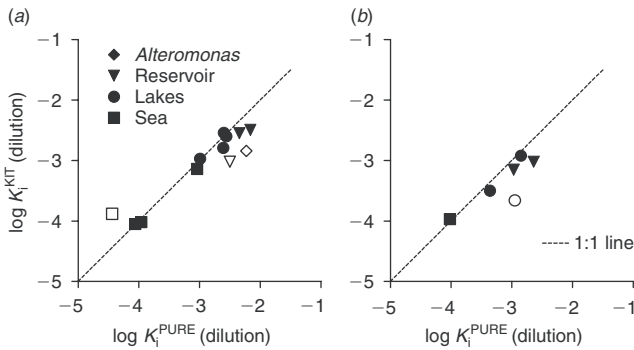


Fig. 2. A comparison of the values of inhibition constants K_i^{KIT} and K_i^{PURE} , obtained in the experiments in which ELFP97 phosphate (purchased from the manufacturer [Molecular Probes] either as a part of the endogenous phosphatase detection kit or as a pure substance) inhibited hydrolysis of 4-methylumbelliferyl phosphate (4MUP) in different aquatic environments (indicated by different symbols). K_i is a measure of enzyme affinity to inhibitor and corresponds to the inhibitor concentration necessary for 50% saturation of the enzymes. Open symbols indicate cases where K_i^{KIT} and K_i^{PURE} differed significantly (F -test; $P < 0.05$). Inhibition constants were estimated using two kinds of inhibition experiments: (a) displacement experiments (see Fig. 1a for experimental design) and (b) saturation experiments (cf. Fig. 1b).

(K_i -ratio = 3.6) and ELFP^{KIT} in 3 cases (K_i -ratios = 0.20–0.31). The geometrical mean of K_i -ratio estimates pooled from displacement and saturation experiments ($n = 19$) was 0.70 (95% confidence interval 0.51–0.95; range 0.20–3.63; median 0.73).

Assuming that the differences in efficiencies between ELFP^{PURE} and ELFP^{KIT} observed in our experiments resulted solely from their different concentrations in the original solutions, we can estimate the original ELFP^{KIT} concentration by dividing the original ELFP^{PURE} concentration (5 mmol L^{-1}) by the mean K_i -ratio obtained in our experiments (0.70). This yields an original ELFP^{KIT} concentration of 7.1 mmol L^{-1} (95% confidence interval 5.3–9.8 mmol L^{-1} ; calculated from the 95% confidence limits of K_i -ratio as above).

Discussion

The inhibition approach employed in the present study proved to be a useful tool by which to gain information about the affinity of extracellular phosphatases for ELFP in natural aquatic systems. Direct quantification of ELFP hydrolysis is complicated by low specific fluorescence of the product of ELFP hydrolysis, ELF alcohol (ELFA), and by the complicated time course of ELFA-fluorescence development (Huang *et al.* 1992; Nedoma *et al.* 2003b; Dignum *et al.* 2004). When applying ELFP as an inhibitor of 4MUP hydrolysis, problems imposed by both low ELFA fluorescence and post-hydrolysis ELFA conversions (Huang *et al.* 1992) are surmounted: enzyme activities as low as $10^{-10} \text{ mol L}^{-1} \text{ h}^{-1}$ are measurable using 4MUP, and any processes following the ELFP hydrolysis do not influence results. In the present study, we successfully applied the inhibition approach in environments covering a broad range of phosphatase activities from 2.6 to $1600 \text{ nmol L}^{-1} \text{ h}^{-1}$.

The fact that ELFP interacts with 4MUP hydrolysis does not strictly mean that it is hydrolysed as well; however, we can

reasonably assume that ELFP hydrolysis took place in all cases in which we observed inhibition of 4MUP hydrolysis. ELFP hydrolysis has been proved by direct microscopic observation of ELFA precipitates in acidified lakes Plešné (Nedoma *et al.* 2003b) and Čertovo (Nedoma and Vrba 2006), in the Římov Reservoir (Štrojsová *et al.* 2003, 2005), and in the oligotrophic West Mediterranean Sea and in the *Alteromonas infernus* cultures (F. Van Wambeke, unpub. data).

The indirect estimates of phosphatase affinity for ELFP in the acidified lakes Plešné ($K_i = 5.2 \text{ } \mu\text{mol L}^{-1}$) and Čertovo ($K_i = 12\text{--}14 \text{ } \mu\text{mol L}^{-1}$) obtained in the present study (Table 1) are close to the direct ELFP affinity estimates obtained at Plešné Lake in saturation experiments ($K_M = 6.5\text{--}16.5 \text{ } \mu\text{mol L}^{-1}$; Nedoma *et al.* 2003b). Thus, it seems that K_i represents a good estimation of K_M for ELFP. However, the lack of direct measurements of K_M for ELFP in other systems (with lower phosphatase activity) does not as yet enable us to draw more general conclusions.

The affinity of extracellular phosphatases for ELFP reflected the trophic status of the system examined. The highest affinity ($K_i \sim 10^{-7} \text{ mol L}^{-1}$) was found at the oligotrophic offshore marine station DYFAMED, the lowest affinity ($K_i \sim 10^{-5} \text{ mol L}^{-1}$) in the eutrophic freshwater reservoir, with intermediate affinity occurring in oligo- to mesotrophic acidified lakes. It is known that bacterial phosphatases display a higher affinity for 4MUP than do algal phosphatases (Chróst and Overbeck 1987; Vrba *et al.* 1993; Hoppe 2003); the observed affinity for ELFP might therefore depend on the composition of the microbial community (with respect to its bacterial and phytoplanktonic components) as well. Our data partly support such a relationship: the highest algal biomass and the lowest ELFP affinity occurred in the Římov Reservoir, while the lowest biomass and the highest affinity were found in marine samples. In contrast, the phosphatase affinity for ELFP in Lake Čertovo was lower than in Plešné Lake (Table 1); however, the former lake is dominated by heterotrophic microorganisms and the latter by algae (Vrba *et al.* 2003).

The properties of extracellular phosphatases of the cultivated marine bacterium *A. infernus* differed significantly from phosphatases in marine samples: in the bacterial culture, K_M for 4MUP hydrolysis was higher by one order of magnitude and K_i for ELFP was higher by two orders of magnitude compared with sea water. This is a rather common phenomenon, because cultivated microorganisms are adapted to higher concentrations of substrates and nutrients used in cultivations. Consequently, affinity constants measured on bacterial cultures should be extended to natural conditions with caution.

Generally, extracellular phosphatases displayed higher affinity for 4MUP than for ELFP: K_M values of 4MUP hydrolysis were on average almost one order of magnitude lower than K_i values of ELFP (Table 1). This is in agreement with the only available direct comparison of 4MUP and ELFP affinities, measured in parallel saturation experiments in Plešné Lake, whereby K_M of 4MUP was three-fold lower than K_M of ELFP (Nedoma *et al.* 2003b). In the present study, the ratio of K_M to K_i fluctuated in the rather broad range between 2 and 18, without a clear trend with respect to the type of locality. The variability in the K_M/K_i ratio was probably partly caused by an experimental error in K_M and K_i determination (variation coefficients for K_M and

K_i estimates were 9–89% and 7–43% respectively). We can also speculate that another source of variability in the K_M/K_i ratio could be connected with different enzymes present in different samples owing to variable plankton composition: different phosphatases may differ in their ability to select between 4MUP and ELFP. The comparatively low ELFP affinity for phosphatases in natural samples should be taken into consideration during planning and interpretation of experiments with ELFP.

The interaction between ELFP and MUP did not deviate considerably from behaviour expected for competitive inhibition. An increase in apparent K_M and no change in V_{max} for 4MUP in the presence of ELFP are predicted for this type of inhibition (Dixon and Webb 1964). In our experiments, both apparent K_M and V_{max} were affected by ELFP; however, while apparent K_M consistently and significantly increased in the presence of ELFP, the effect of ELFP on V_{max} was only weak, irregular (bidirectional) and mostly statistically insignificant. We therefore interpret the interaction of both substrates as competitive.

Our data indicate that all the enzymes cleaving 4MUP are also accessible to ELFP. If the highest ELFP concentration applied was sufficiently above the IC_{50} value observed in the given experiment, then 4MUP hydrolysis was inhibited completely or nearly completely in the presence of this ELFP concentration (92% inhibition on average). The relationship inverse to that observed in the present study – a nearly complete inhibition of ELFP hydrolysis in the presence of a high (1 mmol⁻¹) concentration of 4MUP (measured with image cytometry) – was reported for Římov Reservoir (Štrojsová *et al.* 2003). Both substrates are therefore most probably hydrolysed by the same set of phosphatases.

The original concentrations of ELFP obtained either as pure substance (ELFP^{PURE}; delivered at 5 mmol L⁻¹ concentration) or as part of the endogenous phosphatase detection kit ('Component A', ELFP^{KIT}; delivered at an unspecified concentration) were similar, if not identical. Both ELFP^{PURE} and ELFP^{KIT} exerted, at the same dilution, comparable effects on 4MUP hydrolysis (Fig. 1). More detailed data analysis revealed that, with 95% probability, the concentration of ELFP^{KIT} lies in the interval of 5 to 10 mmol L⁻¹ (concentrations of known components of the detection buffer used in the kit are given by Dignum *et al.* [2004]; pH is 7.5, Molecular Probes, pers. comm.). To date, the lack of knowledge of the ELFP^{KIT} concentration has to a certain extent limited the interpretation of data obtained with ELFP, because the majority of investigators use the kit in their experiments. It follows from our experiments that the assay ELFP concentration in the endogenous phosphatase detection kit, taking into account the 1/20 dilution of the substrate by detection buffer as recommended by manufacturer, is 250–500 μmol L⁻¹. This concentration is about three orders of magnitude higher than our estimates of K_i values for ELFP in oligotrophic marine environments, and is one to two orders of magnitude higher than the same values for freshwater localities (Table 1). Consequently, the ELFP concentration used in the kit should ensure sufficient saturation of phosphatases in natural aquatic environments, specifically >99.5% in marine and 90–99% in freshwater systems. For marine samples, researchers may perhaps consider lowering the ELFP concentration in order to avoid unspecific effects. At any particular locality, it is possible to determine the K_i value for ELFP by applying the inhibition approach suggested

in this study, and to decide on an optimum ELFP dilution to be used in experiments.

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References

- Chang, Y. C., and Prusoff, W. H. (1973). Relationship between the inhibition constant (K_i) and the concentration of inhibitor which causes 50 percent inhibition (IC_{50}) of an enzymatic reaction. *Biochemical Pharmacology* **22**, 3099–3108. doi:10.1016/0006-2952(73)90196-2
- Chróst, R. J., and Overbeck, J. (1987). Kinetics of alkaline phosphatase activity and phosphorus availability for phytoplankton and bacterioplankton in Lake Plußsee (North German eutrophic lake). *Microbial Ecology* **13**, 229–248. doi:10.1007/BF02025000
- Cox, W. G., and Singer, V. L. (1999). A high-resolution, fluorescence-based method for localization of endogenous alkaline phosphatase activity. *The Journal of Histochemistry and Cytochemistry* **47**, 1443–1456.
- Dignum, M., Hoogveld, H. L., Matthijs, H. C. P., Laanbroek, H. J., and Pel, R. (2004). Detecting the phosphate status of phytoplankton by enzyme-labelled fluorescence and flow cytometry. *FEMS Microbiology Ecology* **48**, 29–38. doi:10.1016/J.FEMSEC.2003.12.007
- Dixon, M., and Webb, E. C. (1964). 'Enzymes.' 2nd edn. (Longmans: London.)
- Elser, J. J., Stabler, L. B., and Hassett, R. P. (1995). Nutrient limitation of bacterial-growth and rates of bacterivory in lakes and oceans: a comparative study. *Aquatic Microbial Ecology* **9**, 105–110.
- Gage, M. A., and Gorham, E. (1985). Alkaline phosphatase activity and cellular phosphorus as an index of the phosphorus status of phytoplankton in Minnesota lakes. *Freshwater Biology* **15**, 227–233. doi:10.1111/J.1365-2427.1985.TB00195.X
- González-Gil, S., Keafer, B. A., Jovine, R. V. M., Aguilera, A., Lu, S., and Anderson, D. M. (1998). Detection and quantification of alkaline phosphatase in single cells of phosphorus-starved marine phytoplankton. *Marine Ecology Progress Series* **164**, 21–35.
- Hoppe, H. G. (1983). Significance of exoenzymatic activities in the ecology of brackish water: measurements by means of methylumbelliferyl substrates. *Marine Ecology Progress Series* **11**, 299–308.
- Hoppe, H. G. (2003). Phosphatase activity in the sea. *Hydrobiologia* **493**, 187–200. doi:10.1023/A:1025453918247
- Huang, Z., Terpetschnig, E., You, W., and Haugland, R. P. (1992). 2-(2'-phosphoryloxyphenyl)-4(3H)-quinazolinone derivatives as fluorogenic precipitating substrates of phosphatases. *Analytical Biochemistry* **207**, 32–39. doi:10.1016/0003-2697(92)90495-S
- Jansson, M., Olsson, H., and Pettersson, K. (1988). Phosphatases: origin, characteristics and function in lakes. *Hydrobiologia* **170**, 157–175.
- Jean, N., Boge, G., Jamet, J. L., Richard, S., and Jamet, D. (2003). Seasonal changes in zooplanktonic alkaline phosphatase activity in Toulon Bay (France): the role of Cypris larvae. *Marine Pollution Bulletin* **46**, 346–352. doi:10.1016/S0025-326X(02)00450-2
- Kopáček, J., Hejzlar, J., Borovec, J., Porcal, P., and Kotorová, I. (2000). Phosphorus inactivation by aluminium in the water column and sediments: a process lowering in-lake phosphorus availability in acidified watershed-lake ecosystems. *Limnology and Oceanography* **45**, 212–225.
- Marty, J. C., Chiavérini, J., Pizay, M. D., and Avril, B. (2002). Seasonal and interannual dynamics of nutrients and phytoplankton pigments in the western Mediterranean Sea at the DYFAMED time series station (1991–1999). *Deep-Sea Research. Part II, Topical Studies in Oceanography* **49**, 1965–1985. doi:10.1016/S0967-0645(02)00022-X

- Moutin, T., Thingstad, T. F., Van Wambeke, F., Marie, D., Slawyk, G., Raimbault, P., and Claustre, H. (2002). Does competition for nanomolar phosphate supply explain the predominance of the cyanobacterium *Synechococcus*? *Limnology and Oceanography* **47**, 1562–1567.
- Nedoma, J., and Vrba, J. (2006). Specific activity of cell-surface acid phosphatase in different bacterioplankton morphotypes in an acidified mountain lake. *Environmental Microbiology* **8**, 1271–1279. doi:10.1111/J.1462-2920.2006.01023.X
- Nedoma, J., Padisák, J., and Koschel, R. (2003a). Utilisation of ³²P-labelled nucleotide- and non-nucleotide dissolved organic phosphorus by freshwater plankton. *Archiv fuer Hydrobiologie* **58**, 87–99.
- Nedoma, J., Štrojsová, A., Vrba, J., Komárková, J., and Šimek, K. (2003b). Extracellular phosphatase activity of natural plankton studied with ELF97 phosphate: fluorescence quantification and labelling kinetics. *Environmental Microbiology* **5**, 462–472. doi:10.1046/J.1462-2920.2003.00431.X
- Rengefors, K., Petterson, K., Blenckner, T., and Anderson, D. M. (2001). Species-specific alkaline phosphatase activity in freshwater spring phytoplankton: application of a novel method. *Journal of Plankton Research* **23**, 435–443. doi:10.1093/PLANKT/23.4.435
- Schindler, D. W. (1977). Evolution of phosphorus limitation in lakes. *Science* **195**, 260–262. doi:10.1126/SCIENCE.195.4275.260
- Štrojsová, A., Vrba, J., Nedoma, J., Komárková, J., and Znachor, P. (2003). Seasonal study of extracellular phosphatase expression in the phytoplankton of a eutrophic reservoir. *European Journal of Phycology* **38**, 295–306. doi:10.1080/09670260310001612628
- Štrojsová, A., Vrba, J., Nedoma, J., and Šimek, K. (2005). Extracellular phosphatase activity of freshwater phytoplankton exposed to different *in situ* phosphorus concentrations. *Marine and Freshwater Research* **56**, 417–424. doi:10.1071/MF04283
- Taylor, W. D., and Lean, D. R. S. (1991). Phosphorus pool sizes and fluxes in the epilimnion of a mesotrophic lake. *Canadian Journal of Fisheries and Aquatic Sciences* **48**, 1293–1301.
- Van Wambeke, F., Christaki, U., Giannakourou, A., Moutin, T., and Souvemerzoglou, K. (2002). Longitudinal and vertical trends of bacterial limitation by phosphorus and carbon in the Mediterranean Sea. *Microbial Ecology* **43**, 119–133. doi:10.1007/S00248-001-0038-4
- Vrba, J., Komárková, J., and Vyhnanek, V. (1993). Enhanced activity of alkaline phosphatases – phytoplankton response to epilimnetic phosphorus depletion. *Water Science and Technology* **28**, 15–24.
- Vrba, J., Vyhnanek, V., Hejzlar, J., and Nedoma, J. (1995). Comparison of phosphorus deficiency indices during a spring phytoplankton bloom in a eutrophic reservoir. *Freshwater Biology* **33**, 73–81. doi:10.1111/J.1365-2427.1995.TB00387.X
- Vrba, J., Kopáček, J., Fott, J., Kohout, L., Nedbalová, L., Pražáková, M., Soldán, T., and Schaumburg, J. (2003). Long-term studies (1871–2000) on acidification and recovery of lakes in the Bohemian Forest (central Europe). *Science of the Total Environment* **310**, 73–85. doi:10.1016/S0048-9697(02)00624-1

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